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Linking patterns of freshwater discharge and sources of organic matter within the Río de la Plata estuary and adjacent marshes

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Abstract

We investigated carbon isotopic ratios $\delta^{13}\text{C}$ vs. C/N for surface sediments throughout a large estuarine system (Río de la Plata, RdIP), combined with sediment cores from adjacent marshes to infer main carbon sources. We also evaluated the influence of the El Niño-Southern Oscillation (ENSO) and associated high freshwater discharge events on the organic matter transport within the estuary. The isotopic pattern in surface sediments of the RdIP showed the upper reaches to be influenced by riverine particulate matter ($\delta^{13}\text{C}$ range -24 to -26‰). Similarly, in the sediment cores from marshes of the upper reaches, $\delta^{13}\text{C}$ values decreased from -24‰ in ancient sediments to -28‰ in recent sediments reflecting an increased contribution of organic matter from land including C3 plants and freshwater phytoplankton during the last 50 years. However, the lower reaches represent a depositional environment of marine algae ($\delta^{13}\text{C}$ ranged -21 to -23‰), with no influence of detritus from adjacent marshes indicating minor erosion of the marshes in the lower reaches operating as carbon sink habitats. Our isotopic analysis point that the transport and deposition of terrigenous organic matter within the RdIP and adjacent marsh habitat appear to be both temporally and spatially linked to hydrology patterns.

Additional keywords: Carbon isotopes; C/N ratios; marsh; estuarine sediments; Río de la Plata.

Introduction

Estuaries are characterized by gradients of physical, chemical and biological variables including salinity, nutrients and primary productivity (Whitfield and Elliott 2011). Due to high nutrient loads from river discharge and the diversity of organic matter sources such as zoo/phytoplankton, microphytobenthos, macrophytes, bacteria, estuaries have long been recognized as highly productive habitats (Paerl 2006). Furthermore, the interactions between aquatic and adjacent terrestrial ecosystems through the flow and exchange of organic matter are increasingly apparent (Nakano and Murakami 2001; Cole *et al.* 2011). Such allochthonous resources can significantly subsidize ecosystems of low productivity and influence population dynamics (Polis *et al.* 1997; Pace *et al.* 2004; Abrantes *et al.* 2013), and secondary production (Valiela and Bartholomew 2015). In particular, marsh areas are frequently coupled to adjacent aquatic ecosystems, such as estuaries, by contributing with organic matter exports (Abrantes and Sheaves 2008; Whitfield and Elliott 2011). Therefore, the import of terrestrial organic matter into estuarine systems is critical to reliably assess both ecosystem functioning and productivity (Abrantes *et al.* 2013).

The continental freshwater input of large estuaries into the ocean is commonly modulated by climatic forcing (Piola *et al.* 2008; Barreiro 2010; García-Rodríguez *et al.* 2014). In particular, an important characteristic of aquatic systems in South America is the variability associated with the El Niño-Southern-Oscillation (ENSO) events (a recurring 2–7 year cycle of coupled ocean–atmosphere dynamics in the equatorial Pacific) and rainfall patterns (Mechoso and Perez-Iribarren 1992; Depetris *et al.* 1996; Barreiro 2010). Strong ENSO influence occurs during warm-phase ENSO events that are associated with increased rainfall events over southeastern

South America (Barreiro, 2010), leading to high discharge events in the Uruguay/Paraná River (Bischoff *et al.* 2000; Camilloni and Barros 2000). In estuarine ecosystems, such periods of high discharge can considerably modify the concentrations of material and nutrients transferred between ecosystems (Goñi *et al.* 2009; Valiela and Bartholomew 2015). Hence, studying both modern and historical sources of organic matter from the continental and fluvial to the transitional estuarine region and the adjacent ocean, is important to infer the transport and budgets of organic matter in relation to hydrological and oceanographic processes (e.g. Burone *et al.* 2013).

Stable carbon isotope composition ($\delta^{13}\text{C}$) together with C/N ratios of bulk sedimentary organic matter have been successfully used to infer the relative influence of marine vs. terrestrial sources in coastal environments (e.g. Müller and Voss 1999; Malamud-Roam and Ingram 2001; Tue *et al.* 2011; Remeikaitė-Nikienė *et al.* 2016). This is because $\delta^{13}\text{C}$ values are usually minimally affected by diagenetic processes in sediment organic matter (Thornton and McManus 1994; Lautenschlager *et al.* 2014). Terrestrial material from C3 pathways exhibits $\delta^{13}\text{C}$ values that range between -32 and -21‰. On the other hand, C4 plants typically show $\delta^{13}\text{C}$ values between -17 to -9‰, while marine organic matter typically displays $\delta^{13}\text{C}$ values between -24 to -18‰ (Meyers 1994; Lamb *et al.* 2006). Furthermore, terrestrial vegetation has relatively high C/N ratios, usually higher than 12, due to the lignin and cellulose abundance, but C/N ratios from 4 to 10 are characteristic for marine organic substances (Meyers 1994, 2003). In this context, distinctive suits of C/N and $\delta^{13}\text{C}$ values characteristics of C3, C4 plants, marine, and freshwater algae can be used to identify major sources of organic matter in coastal sediments (Lamb *et al.* 2006; Tue *et al.* 2011; Remeikaitė-Nikienė *et al.* 2016).

93
94 Some studies (Botto *et al.* 2011; Burone *et al.* 2013) have applied stable isotope analysis in
95 surface sediments within the Río de la Plata (RdIP) answering the question of the origin of
96 organic matter (i.e. terrigenous or marine), but investigations on land-water coupling including
97 the adjacent marshes of the RdIP are less common. Therefore, the aim of this study was to gain
98 insight into the sources and distribution of organic matter within the RdIP to infer both the
99 mechanisms and factors responsible for variations in organic matter material fluxes. We assessed
100 $\delta^{13}\text{C}$ values and C/N ratios in both sediment cores and surface sediment samples along a salinity
101 gradient to assess the spatial patterns in the sources of organic matter (from riverine and
102 terrestrial plant material to marine phytoplankton). In addition, we explored the temporal pattern
103 related to river discharge variability associated with ENSO events. We hypothesized that
104 different patterns of freshwater discharge associated with ENSO events induce a change in the
105 sources of organic matter deposited within the RdIP estuary and the fringing marshes.

106

107 **Materials and methods**

108

109 *Study area*

110 The study area comprises the inner, middle and outer regions of the RdIP estuary (34°10'–36°20'
111 S and 55°00'–58°30' W) and the adjacent marsh habitats (Fig. 1). This estuary region represents
112 a highly productive area that sustains valuable artisanal and coastal fisheries in Uruguay and
113 Argentina (FREPLATA 2004). The RdIP estuary is one of the world's largest estuaries, with an
114 average annual river discharge of ca. 24,000 m³ s⁻¹ representing a confluence of the Paraná and
115 Uruguay River with a mixohaline area of 38,000 km² (Mianzan *et al.* 2001). The estuary is

approximately 320 km in length and 250 km in width (minimum of 40 km). It is relatively shallow with only 5-15 m of water depth, and the tide range is small, averaging about 0.60 m (Framiñan and Brown, 1996). River discharge increases from April to June with a sharp rise in May. It reaches minimum values from December to March (Guerrero *et al.* 1997). Salinity varies significantly during the year in negative correlation with river runoff; lowest salinities are found during periods of high river discharge from April to August, and highest salinities are observed in December and January (Guerrero *et al.* 1997). The estuary can be divided into three sections based on its morphology, and the prevailing hydrodynamics (Framiñan and Brown 1996; Nagy *et al.* 2002). The upper reaches are characterized by freshwater conditions and water column depths lower than 5 m (Mianzan *et al.* 2001; Nagy *et al.* 2002). The middle reaches are characterized by the salt intrusion boundary, prevailing shallow waters and intermediate channels (<10 m). A submersed bar, called Barra del Indio shoal, is a sedimentary facies consequent of the bottom salinity front and represent the boundary between the riverine regime and the salt wedge regime, following approximately, the 10 m isobath (Guerrero *et al.* 1997; Fig. 1). The lower reaches with maritime channel (10-20 m depth), reflect a gradual increase in the marine influence including a small transitional section together with the marine part of the estuary with a characteristic salinity and seston gradient (Calliari *et al.* 2005). Here, the Subtropical Convergence emerges where the southward-flowing Brazil Current meets the northward-flowing Malvinas (Falkland) Current, resulting in complex frontal systems and hydrographic structures like eddies and meanders (Calliari *et al.* 2009). The average annual suspended sediment load from the Paraná and Uruguay rivers is 79.8×10^6 tons yr^{-1} , and is characterized by 75% coarse to medium silt, 15% fine to very fine silt, and 10% clay (Giberto *et al.* 2004).

Large parts of the estuary are fringed by freshwater and salt marshes, whereas the salt marshes along the Uruguayan coast are dominated by halophytes including *Spartina montevidensis* (C4), *S. longispina* (C4), *Juncus acutus* (C3) and *Salicornia ambigua* (Mianzan *et al.* 2001). Along the Argentinean coast, marshes of the Samborombón bay are dominated by the cordgrasses *Spartina alterniflora* (C4), *Spartina densiflora* (C4) and *Sarcocornia perennis* (C3) (Isacch *et al.* 2006). Meanwhile, the freshwater marshes in the inner estuary are dominated by small emergent plants including *Ludwigia* spp., *Alternanthera philoxeroides*, *Echinodorus* sp., *Eryngium* sp.

Data collection

Monthly average values of the Paraná and Uruguay rivers discharge were obtained from the Integrated Hydrologic Database (<http://bdhi.hidricosargentina.gov.ar/sitioweb/frmInicial.aspx>) by the Secretary of Water Resources, Argentina. Stations representing the Paraná and the Uruguay River are “Túnel subfluvial” and “Paso de los Libres”, respectively. Relative anomalies of river discharge are shown in Fig. 2, calculated following Piovano *et al.* (2004) as: $Anomaly\ Q = (Q_i - Q_{average}) / Q_{average}$, where Q_i is the sum of the monthly discharge averages per year and $Q_{average}$ is the average of all values of the time series (i.e. 1909 AD-present). In specific years such as 1982/1983, and 1997/1998 the Paraná and Uruguay rivers experienced high river discharges, which were greater than two times the annual mean of $24,000\ m^3\ s^{-1}$. Furthermore, during the period 2000/2001 only the Uruguayan river showed increased mean river discharge, and during the years 1991/1992 only the Paraná river discharges were twice its average (Fig. 2). These anomalies coincide with strong El Niño events according to the Oceanic Niño Index (ONI, http://www.cpc.ncep.noaa.gov/products/analysis_monitoring/ensostuff/ensoyears.shtml). The

ONI is defined as a three month running mean of sea surface temperature (SST) anomalies in the Niño 3.4 region (5°N–5°S latitude, 120°W–170°W longitude), with values greater than +(-)1.5°C indicating strong El Niño (La Niña) conditions.

Sampling and stable isotope analysis

The data of six sediment cores that were collected in marsh sites adjacent to the RdIP by Schuerch *et al.* (2016) were used for the present analysis (Fig. 1). Coring locations were in the inter- and supratidal zones of the marshes (above the mean high water level) and characterized by dense vegetation cover, in order to avoid a stress caused by currents and waves. Except for core 08-1, all cores were collected on the river banks of the RdIP. Coring site 08-1 was located about 900 m from the RdIP in the river mouth of the Río Salado. Cores were extracted using PVC tubes with an inner diameter of 10.3 cm. Core lengths ranged from 51 to 115 cm, averaging 91.5 cm. Cores were sliced into layers of 2 cm for all depths between 0 and 20 cm, 3 cm for depths between 20 and 50 cm and 5 cm for all layers below (Schuerch *et al.* 2016). For core 12-1 and 10-1 a total of 21 and 24 sediment samples were collected respectively. A total of 21 sediment samples were collected from core 19-1, 26 samples for the core 09-1, 25 samples for the core 08-1, and 22 for the core 02-1.

The sediment samples were dried at 60°C until constant weight and manually ground to a fine homogeneous powder using mortar and pestle. Elemental and bulk isotope analysis was conducted by the element analyzer “NA1110” (CE Instruments, Rodano, Italy) and the mass spectrometer “Delta Plus” coupled via a “Conflo III” (Werner *et al.* 1999).

185 Surface sediments samples were collected from 25 stations within the RdIP estuary during 1-2
186 weeks cruises (FREPLATA-Ifremer research) in March (warm season) and June (cold season)
187 2010. Sampling stations were located in the upper, middle and lower reaches of the estuary (Fig.
188 1). Sediments samples were collected using a Van Veen grab. This approach has also been used
189 in previous studies of biochemical composition in surficial sediments (e.g. García-Rodríguez *et*
190 *al.* 2014). In the laboratory, samples were acid washed (3 h in 10% HCl) to remove carbonates,
191 rinsed in distilled water and then dried again for 24 h at 60 °C. All sediment samples were
192 ground to a fine homogeneous powder using mortar and pestle for elemental (C_{org} , N_{total}) and
193 bulk isotope analyses, including the measurement of $\delta^{13}C$ ratios. All samples were then analyzed
194 using a Thermo Finnigan Flash EA 1112 series elemental analyzer equipped with a Thermo
195 Finnigan Delta Plus XP isotope ratio mass spectrometer by the *Centro de Aplicaciones de*
196 *Tecnología Nuclear en Agricultura Sostenible*, University of Uruguay.

197

198 Spatial changes of the $\delta^{13}C$ and C/N ratios for the marsh cores and surface sediments were
199 determined by non-metric multi-dimensional scaling plots (nMDS: Kruskal and Wish 1978)
200 using the software Paleontological Statistics (PAST) 1.42 (Hammer *et al.* 2001). For this
201 purpose, similarity matrices were constructed on normalized data (by subtracting the mean from
202 each datapoint and dividing by the standard deviation) using the Euclidean distance similarity
203 measure. Analyses of similarities (ANOSIM with 1000 permutation) were performed on each
204 Euclidean similarity matrix to determine resemblance in $\delta^{13}C$ and C/N ratios among sites within
205 the RdIP and adjacent marsh sites. For the marsh cores we included all dates/depths available,
206 and for the superficial estuary sediment data of both sampling periods (March and June) were
207 included into this analysis.

Sample dating

Radionuclide analysis (^{210}Pb and ^{137}Cs) was conducted by Schuerch *et al.* (2016) in order to determine the years of deposition for each of the sediment layers. Alpha spectrometry, using an *Ortec Octète Plus* alpha-spectrometer, was applied to determine sediment ages of the cores 02-1, 12-1 and 19-1. Prior to the analysis, the sediment samples were digested in the presence of ^{209}Po yield tracer and by means of 3 ml HF, 5 ml HNO_3 , 2 ml HCL. Digestion was accelerated by heating the samples with a *MARS 6* microwave system, followed by the spontaneous deposition of polonium on silver disks and the measurement of the resulting ^{209}Po activity. Reference samples for validation purposes were conducted using uranium reference material *UREM-11*.

Gamma spectrometry, using high-purity germanium detectors (*CANBERRA BE3830P*), was employed for age-dating of the cores 08-1 and 10-1 (for details see Schuerch *et al.* 2016). Aliquots of 10-15 g were, therefore, placed in sealed vials and allowed to sit for three weeks until equilibrium has established between ^{226}Ra and ^{214}Bi . ^{210}Pb , ^{214}Pb , ^{214}Bi and ^{137}Cs were measured via their gamma peaks at 45.6 KeV, 352 KeV, 609 KeV, and 661 KeV, respectively. Unsupported ^{210}Pb , needed for dating purposes, was calculated as the difference between the total ^{210}Pb and the ^{226}Ra intensity. ^{226}Ra , in turn, was calculated as the mean value of ^{214}Bi and ^{214}Pb , both of which were assumed to be in equilibrium with ^{226}Ra .

Based on the unsupported ^{210}Pb activities (see supplementary material), sediment ages were derived by means of the *Constant-Flux* (CF), or *Constant-rate of supply* (CRS), dating model in combination with the *Constant Flux Constant Sedimentation* (CFCS) model, where the captured

231 ^{210}Pb inventory was incomplete (Appleby and Oldfield 1978; Sanchez-Cabeza and Ruiz-
232 Fernández 2012; Schuerch *et al.* 2016).

233

234 **Results**

235

236 *$\delta^{13}\text{C}$, and C/N from marsh sites*

237 The $\delta^{13}\text{C}$ ratios for the cores 12-1 and 10-1, located in the Paraná Delta near the freshwater
238 discharge in the upper reaches, showed a considerable range of $\delta^{13}\text{C}$ values between -23.2 and -
239 28.3‰ (Fig. 3). Downcore sediment samples from the two cores in the upper reaches of the
240 estuary displayed higher $\delta^{13}\text{C}$ ratios with an upward-decreasing trend that is more pronounced in
241 core 10-1 as compared to core 12-1. Sediments showed minimum $\delta^{13}\text{C}$ ratios of -28.3‰ in core
242 12-1 (in a depth of 5 cm) and -27.7‰ in core 10-1 (in a depth of 5 cm). In sediments from core
243 10-1, $\delta^{13}\text{C}$ values particularly sharply decreased during the extreme ENSO events of 1982/1983,
244 1997/1998 and 2000/2001. In the core 12-1, $\delta^{13}\text{C}$ decreased abruptly during 2000/2001 below -
245 27‰. Furthermore, C/N ratios of sedimentary organic matter in the upper reaches of the estuary
246 were generally higher than 9, and ranged between 9 and 13 (Fig. 3). The C/N ratios in core 10-1
247 increased abruptly above 10 during the extreme ENSO event of 1997/1998, whereas in core 12-
248 1, located in vicinity of Uruguay river, the amplitude of the ENSO related peaks in C/N ratios are
249 less pronounced (e.g. the 2000/2001 ENSO event). Similar to the abrupt $\delta^{13}\text{C}$ decrease after the
250 2000/2001 ENSO, the C/N ratios in core 12-1 show a remarkable drop.

251

252 Over the length of the four cores from the lower reaches of the estuary (cores 19-1, 08-1, 09-1
253 and 02-1), $\delta^{13}\text{C}$ values showed a range of 8.6‰, i.e. between -10.4 and -19.0‰ (Fig. 4a, b, e). In

sediments from the core located furthest away from the freshwater discharge (core 19-1), $\delta^{13}\text{C}$ showed minimum values in the years 1986/1987, 1993/1994, and 1997/1998 with -17.9, -17.5, and -17.6‰ respectively. The C/N ratios equal to zero in the upper section of core 19-1 can be explained as a consequence of the possible loss of organic carbon derived from decomposition of organic matter and remineralization that introduces uncertainty to infer organic matter (Müller 1977; Meyers 1997; Sampei and Matsumoto 2001). $\delta^{13}\text{C}$ values in sediments from core 08-1 in the lower reaches decrease abruptly after the ENSO events of 1982/1983 and 1992/1993 below -18‰. Simultaneously, C/N ratios drop below 12 and 10 respectively (Fig. 4d). For the core 09-1 isotopic values of $\delta^{13}\text{C}$ reach their minimum of -17.8 and -17.6‰ in sediments at 70 and 50 cm depth respectively (Fig. 4e). For this core, no dating is available hence isotopic patterns are analyzed as a function of depth only (Fig. 4e,g). Maximum $\delta^{13}\text{C}$ values together with highest C/N ratios of -9‰ and 22, respectively were detected in sediments in about 6 cm depth. In the marsh core 02-1, located in the lower reaches of the RdIP along the Uruguayan coast, $\delta^{13}\text{C}$ ranged between -19.5 and -16.8‰ and tended to increase from 1960 to present days (Fig. 4f).

Analysis of the marsh cores from the upper reaches (core 12-1 and 10-1) indicated that organic matter is composed of a mixture of freshwater algae and terrestrial C3 plants (core 10-1: range of $\delta^{13}\text{C}$ between -25.5 and -27.6‰, range of C/N between 10.4 and 12.3; core 12-1: range of $\delta^{13}\text{C}$ between -28.5 and -25.6‰, range of C/N ratio between 9.6 and 13.3, Fig. 5). On the other hand, $\delta^{13}\text{C}$ and C/N values of sediment cores from saltmarshes located in the lower reaches were associated with a mixture of marine algae and C4 marsh grasses (Fig. 5).

$\delta^{13}\text{C}$, and C/N of surface samples within the RdIP

The $\delta^{13}\text{C}$ values for surface sediments in collected during the cold season showed a narrower range than the $\delta^{13}\text{C}$ values of the samples collected during the warm season (-20.2 to -24.6‰ and -21.2 to 26.2‰ respectively). $\delta^{13}\text{C}$ values showed a clear trend along the estuarine gradient with sediments from the upper reaches having the most depleted carbon (-25 to -26‰; Fig. 6), while such values progressively increased towards the middle and lower reaches (Fig. 6). The C/N ratios in surface sediments within the RdIP showed little difference between seasons with values ranging from 6.6 to 8.9 for samples collected in March and from 6.5 to 8.9 in for the samples collected in June. The variability of C/N ratios among the sites also appears to be small (Fig. 6).

Based on the $\delta^{13}\text{C}$ values and C/N ratios, the marsh core data were grouped in nMDS plots (all depths pooled Fig. 7a), while being separated from surface sediment samples from the upper, middle and lower reaches of the RdIP (Fig. 7b), . In both cases samples from the middle reaches spread between the upper and lower reaches groups. Not only the stress values (< 0.1) indicated a significant two-dimensional representation of the groups, but also the ANOSIM results revealed significant differences in the isotopic composition among all surface samples within the RdIP ($R = 0.83$, $P = 0.0001$) and among the different marsh sites ($R = 0.58$, $P = 0.0001$). Furthermore, the nMDS plot on $\delta^{13}\text{C}$ values and C/N ratios including all sediment samples (core data and surface sediments) showed different distribution patterns and a clear separation between marsh and estuarine sites (Fig. 7c).

Discussion

Our isotopic data highlighted differential composition and organic matter along the RdIP and adjacent marshes, suggesting different sources. The isotopic measurements in marsh sediments from the upper reaches are consistent with allochthonous organic carbon derived from freshwater sources and terrestrial C3 plants while marsh sediments located in the lower reaches are consistent with a high accumulation of marine phytoplankton and C4 marsh grasses (e.g. *Spartina* spp.). In the surface sediments of the RdIP, lower $\delta^{13}\text{C}$ values towards the upper reaches also indicated increasing deposition of terrestrial C3 plants, while towards the lower reaches the deposition events of marine algae would be more significant, with no carbon contribution of marsh grasses. Such a distributional pattern can clearly be related to the salinity gradient observed from the upper reaches to the middle estuary and the marine domain (Fig. 8), with a probable stronger influence of dissolved inorganic carbonates from freshwater which are characterized by low $\delta^{13}\text{C}$ in lower salinities (Fry and Sherr 1984). These results are consistent with previous studies suggesting that terrestrial organic matter is available as energy source for benthic consumers within the upper and middle reaches in the RdIP (Botto *et al.* 2011; Burone *et al.* 2013; Venturini *et al.* 2014; Marrero *et al.* 2014). Furthermore, our results are in agreement with recent findings showing that food sources including both estuarine microphytobenthos and diatoms show lower $\delta^{13}\text{C}$ values at low salinities rather than at high salinities (Lebreton *et al.* 2015).

Organic carbon sources in marsh core sediments

In our study, the $\delta^{13}\text{C}$ profiles obtained from the two marsh cores near the freshwater discharge (10-1 and 12-1) showed a decreasing trend from -24 to -28‰. Such depleted $\delta^{13}\text{C}$ values are indicative of a predominant freshwater source influence (with a typical range of -30 to -26‰,

Meyers 1994). Consistently, previous studies also attributed lower $\delta^{13}\text{C}$ values to stronger influence of riverine particulate matter mostly composed of freshwater phytoplankton and C3 plant material with $\delta^{13}\text{C}$ values consistently between -29 and -26‰, that corresponded well with those of the marsh cores from upper reaches of the RdIP (e.g. Wilson et al. 2005; Tue *et al.* 2011; Lebreton *et al.* 2015). However, in the sediment cores from marshes adjacent to the lower reaches of the RdIP (02-1, 08-1, 09-1, and 19-1) $\delta^{13}\text{C}$ values with a range from -9 to -19.5‰ reflect a considerably different composition of the organic matter. The remarkably low C/N values in core 08-1 (<10) from the lower reaches confirmed a contribution of algal marine sources in the sedimentary organic matter (Lamb *et al.* 2006). In addition, the $\delta^{13}\text{C}$ values of *Spartina* spp. that dominates in marshes adjacent to the lower reaches of the RdIP showed values around -11‰ (Botto *et al.* 2011), and therefore fall within the range of our $\delta^{13}\text{C}$ values registered in marsh sediments from the lower reaches suggesting that C4 marsh grasses contribute to the deposited sediment (autochthonous sediment deposition) together with marine organic matter (allochthonous sediment deposition).

Organic carbon sources in surface sediments within the RdIP

In surface sediments along the RdIP, the progressive increase in $\delta^{13}\text{C}$ from the upper towards lower reaches, with average values around -21‰, indicated a greater influence of ^{13}C -enriched sources in the lower reaches of the estuary, including marine phytoplankton, and dissolved inorganic carbon (Chanton and Lewis 1999). Furthermore, sedimentary C/N ratios were closer to the ratios for marine phytoplankton, characterized by higher nitrogen-containing compounds (C/N around 6-10; Meyers 1994; Lamb *et al.* 2006), thereby suggesting a lower contribution of terrestrial derived organic matter in sediments from the lower reaches of the RdIP. This isotopic

pattern is opposed to the $\delta^{13}\text{C}$ values of -26 to -25‰ (i.e. in the range of terrestrial plant material and freshwater algae) found in the surface sediments from the upper reaches, indicating a contribution of terrestrial organic matter, and similar to what we found in the freshwater marshes. These results suggest that within the upper reaches of the RdIP river transports of organic matter are significantly contributing terrestrial derived material, including river POM, that represent a mixture of detritus and freshwater algae. This finding is consistent with the observation of Burone *et al.* (2013) who report an increased seaward marine influence in the transition zone of the RdIP. Previous studies in other estuaries also found a decreasing trend of $\delta^{13}\text{C}$ values in sediments from the lower reaches towards the upper reaches (Middelburg and Nieuwenhuize 1998; Zimmerman and Canuel 2001; Lautenschlager *et al.* 2014), indicating a higher contribution of ^{13}C -enriched marine sources to the allochthonous terrestrial sources towards the sea. The nMDS ordination analysis for isotope data and C/N ratios consistently grouped together samples from upper and lower reaches, while those of the middle reaches were dispersed into both groups. Furthermore, marsh vegetation seems not to influence adjacent coastal environments within the lower reaches of the RdIP dominated by marine derived organic matter. These results imply that most of the C4 marsh grasses (e.g. *Spartina* spp.) were processed within marsh habitats, and therefore act as a sink for organic carbon including marine phytoplankton.

Contribution of ENSO events to organic matter in core sediments

ENSO events influence the organic matter sources along the RdIP. Our $\delta^{13}\text{C}$ analysis in sediments from core 10-1, located in vicinity of the Paraná River, showed a generally decreasing trend towards present days, concurrent with an increase in the river discharge in the Paraná River

368 since the 1970s (Berbery *et al.* 2006; Marrero *et al.* 2014). Furthermore, in the sediment core 12-
369 1, affected by the Uruguay River, abrupt changes in $\delta^{13}\text{C}$ values occurred in 2000/2001. This
370 decreasing trend in $\delta^{13}\text{C}$ values together with C/N ratios showing distinct peaks higher than 10
371 during the strong ENSO events in the deposited organic matter supports the hypothesis that high
372 periods of freshwater discharge that evolved during warm-phase ENSO events lead to an
373 increased contribution of freshwater sources and terrestrial plant material. These results agree
374 with findings of other studies that found evidence that the patterns of freshwater discharge and
375 marine water intrusion moving in opposite direction control the dynamic energy of terrestrial
376 subsidy and deposition to adjacent estuarine environments (Hoffman and Bronk 2006; Goñi *et al.*
377 2009; Hoeninghaus *et al.* 2011). Carbon sources from terrestrial organic matter, including
378 invertebrates and autumn leaf fall, could have greater importance for aquatic ecosystems
379 subsidizing aquatic organisms during high discharge periods (Junk *et al.* 1989). Furthermore,
380 previous articles highlighted the important influence of ENSO events and river runoff patterns on
381 fish assemblage (Garcia *et al.* 2001), on shrimp catches in the Patos Lagoon in southern Brazil
382 (Moller *et al.* 2009), and also in ecosystem properties of the RdIP (Vögler *et al.* 2015). Our
383 results (together with those of Schuerch *et al.* 2016) highlighted the effects of interannual climate
384 variability linked to ENSO events, on the influence of the terrestrial/marine organic matter inputs
385 drained into the RdIP estuary. Such a variability in organic matter composition may have
386 significant implications on the production of species and food web structures. Furthermore, as a
387 result of the higher inputs of nutrients and terrestrial organic matter during periods of high
388 freshwater discharge, phytoplankton production may increase within the estuary, as reported by
389 Machado *et al.* (2013), together with changes in the structure of the phytoplankton in coastal
390 areas of the RdIP (Sathicq *et al.* 2015). In addition to the pattern of freshwater discharge, we

identified different variability of $\delta^{13}\text{C}$ values and C/N ratios among cores, which implied that the influence of marine vs. terrestrial organic matter was site specific. Local differences are likely caused by spatial difference in morphology, elevation, local hydrodynamics as well as the distance of the coring sites from the main channel, which impact the vegetation patterns, sources of organic matter and hence the variability in $\delta^{13}\text{C}$ values and C/N ratios (Isacch *et al.* 2006; Hoeinghaus *et al.* 2011). Furthermore, differences among cores may arise due to the inherent uncertainties (ca. ± 3 years) of the radiocarbon dating (Schuerch *et al.* 2016).

Conclusions

Based on $\delta^{13}\text{C}$ values and C/N ratios we found that sources of organic matter in sediments within the RdIP can be separated into the upper, middle, and lower reaches. The RdIP estuary and adjacent marsh sites in the upper reaches received organic matter mostly from allochthonous sources (freshwater sources and C3 plant detritus), and autochthonous carbon sources tend to increase towards the lower reaches (marine algae). Additionally, our data suggested that marsh habitats did not supply sediments into the RdIP estuary, indicating minor erosion of marsh in the lower RdIP reaches. The marshes, therefore appear to act as a sediment sink, which has important implications for the marsh's function as sink for pollutants and nutrients in the estuary. Furthermore, extreme ENSO events tied with high freshwater discharge appear to play an important role in linking the terrestrial derived sources from adjacent habitats with the estuarine ecosystems. The information presented here is important to help infer and interpret similar processes of organic matter transport and composition in other large estuarine systems with distinct changes in continental terrigenous inputs.

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663

FIGURE CAPTIONS

Fig. 1. Map of the RdIP estuary and location of all sample sites. The station codes describe the name of each site. Black stars indicate the location of the six marsh cores. Triangles and circles indicate the location of the surface samples along the estuary with 24 sampling stations; white circles: upper reaches; black triangles: middle reaches; and grey circles: lower reaches. Dashed line represent a submersed bar (Barra del Indio shoal). Study period from March to June 2010.

Fig. 2. Standardized discharge in the Paraná and Uruguay rivers below and above the mean estimated annual river discharge (m^3/s). Strong ENSO events (ONI indices > 1.5) are indicated using triangles.

Fig. 3. $\delta^{13}\text{C}$ values and C/N ratios of sedimented organic matter versus age at two different saltmarsh areas located in the upper reaches of the RdIP estuary (cores 12-1 and 10-1). ENSO events are indicated using triangles on the y-axes. Core sites are indicated in Fig. 1.

Fig. 4. $\delta^{13}\text{C}$ values and C/N ratios of sedimented organic matter versus age at four different saltmarsh areas located in the lower reaches of the RdIP estuary (cores 19-1, 08-1, 09-1 and 02-1). ENSO events are indicated using triangles on the y-axes. Note that the y-axis of the plot for the core 09-1 indicates depth in the sediments of the marsh core. Note also the different y-axis scales. Core sites are indicated in Figure 1 Cores 19-1 and 02-1 presented no clear pattern in N

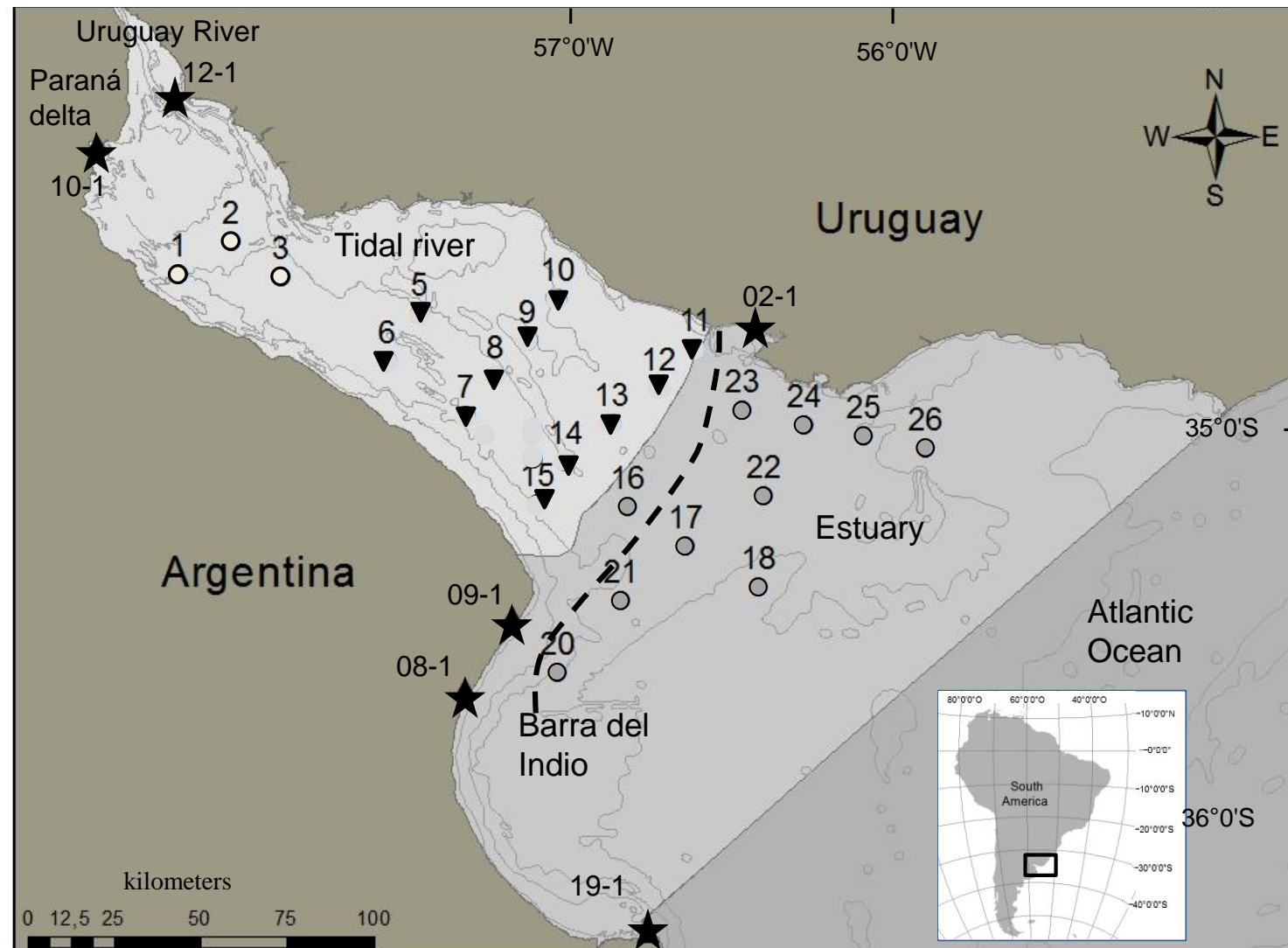
values that contribute with imprecision to infer organic matter sources; consequently C/N data was judge unreliable and not further considered.

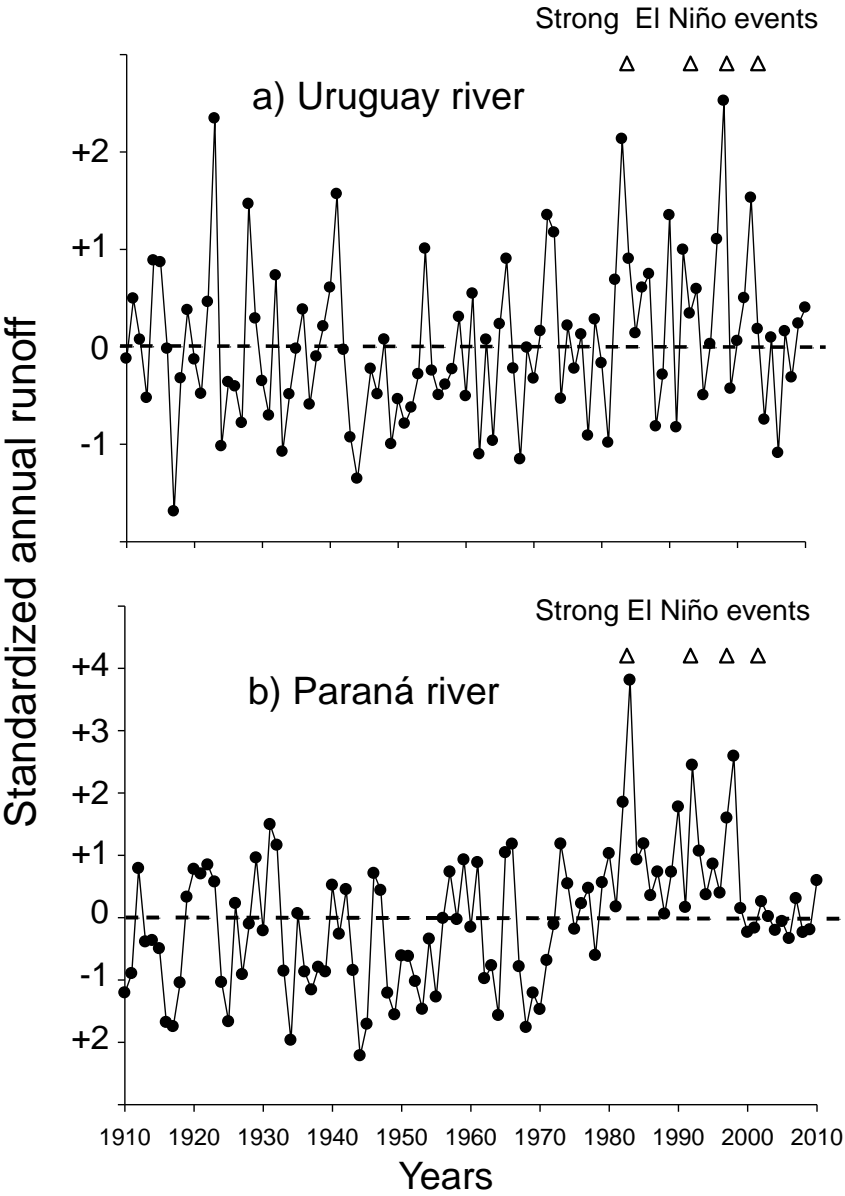
Fig. 5. Relationship between $\delta^{13}\text{C}$ and C/N values of the sediments from the saltmarshes habitats located in the upper (10-1 and 12-1) and lower reaches (09-1, 08-1, 19-1 and 02-1) of the RdIP, including typical ranges of sources according to data presented by Meyers (1994) and Lamb *et al.* (2006). Core sites are indicated in Figure 1. Cores 19-1 and 02-1 presented no clear pattern in N values that contribute with imprecision to infer organic matter sources; consequently C/N data was judge unreliable and not further considered.

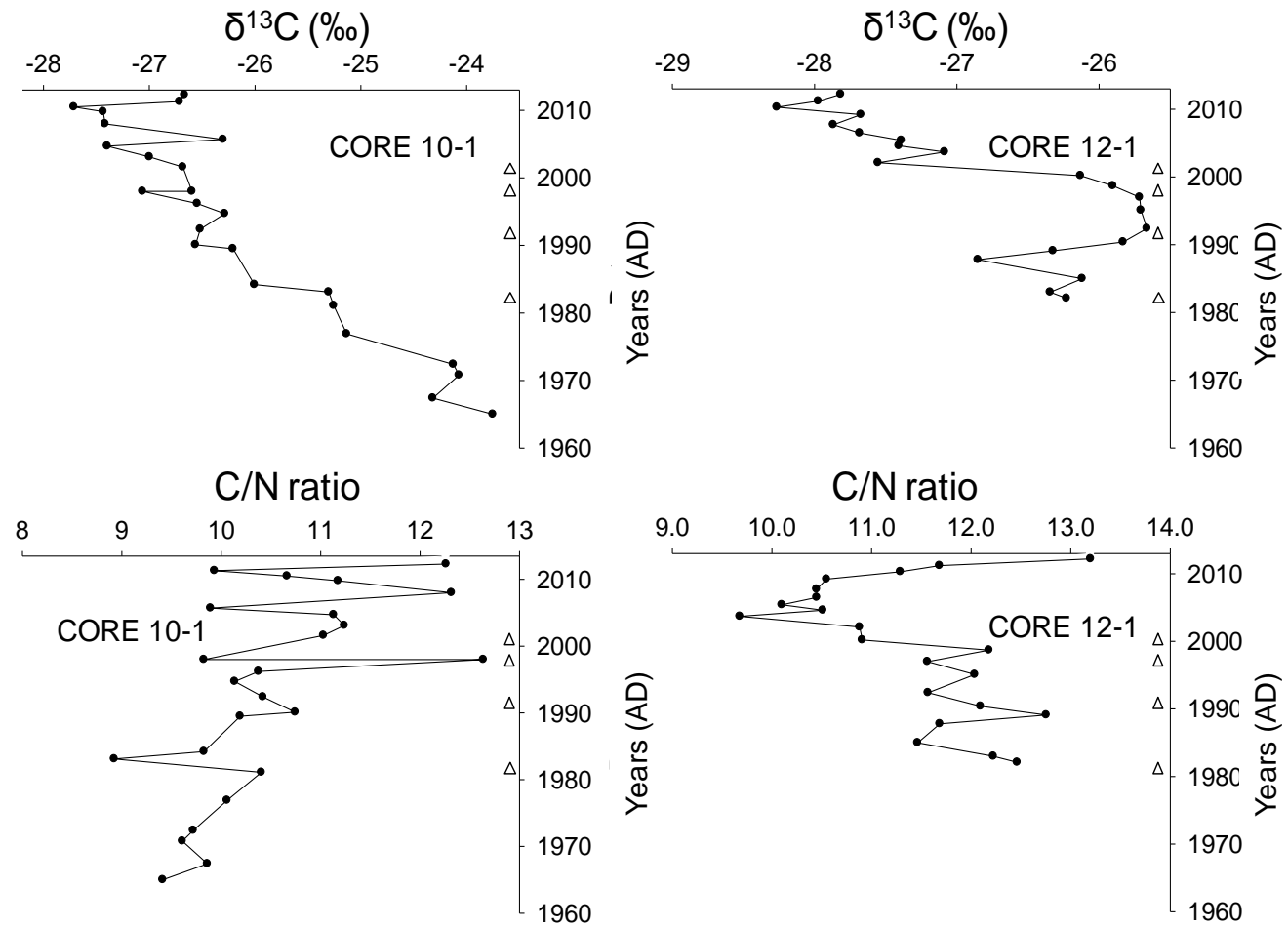
Fig. 6. Relationship between $\delta^{13}\text{C}$ and C/N values of the surface sediments from each sampling station within the RdIP estuary (upper, middle and lower reaches) in March (warm season) and June (cold season) 2010, including typical ranges of sources according to data presented by Meyers (1994) and Lamb *et al.* (2006).

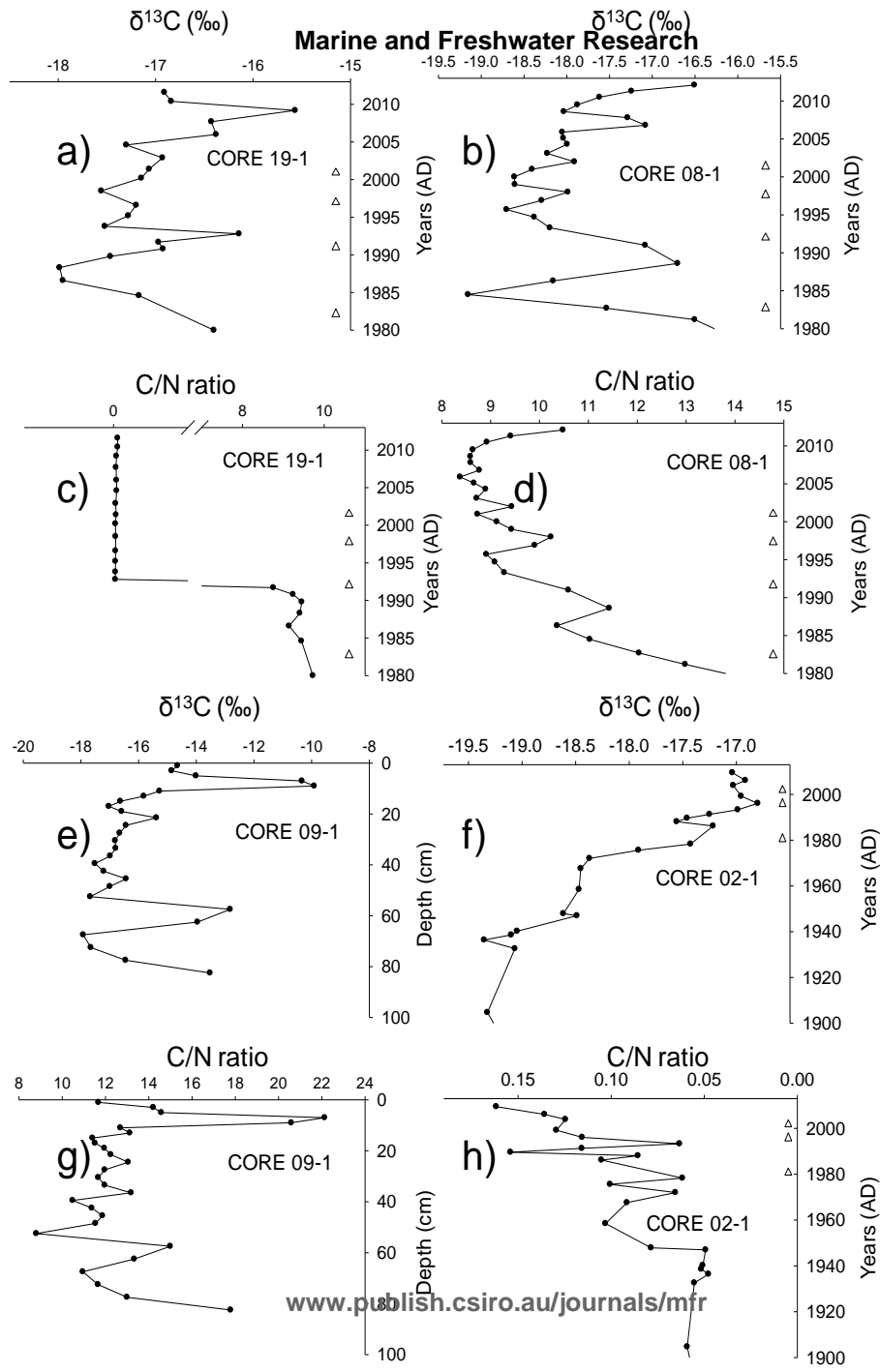
Fig. 7. Two-dimensional nMDS ordination of the C/N ratios and isotopic composition of sediments from the upper (U), middle (M) and lower reaches (L) within the RdIP estuary (a) and adjacent marsh sites (b). The data are standardized and the matrix was performed using normalized Euclidian distances. For the marsh cores we include all dates in the analysis.

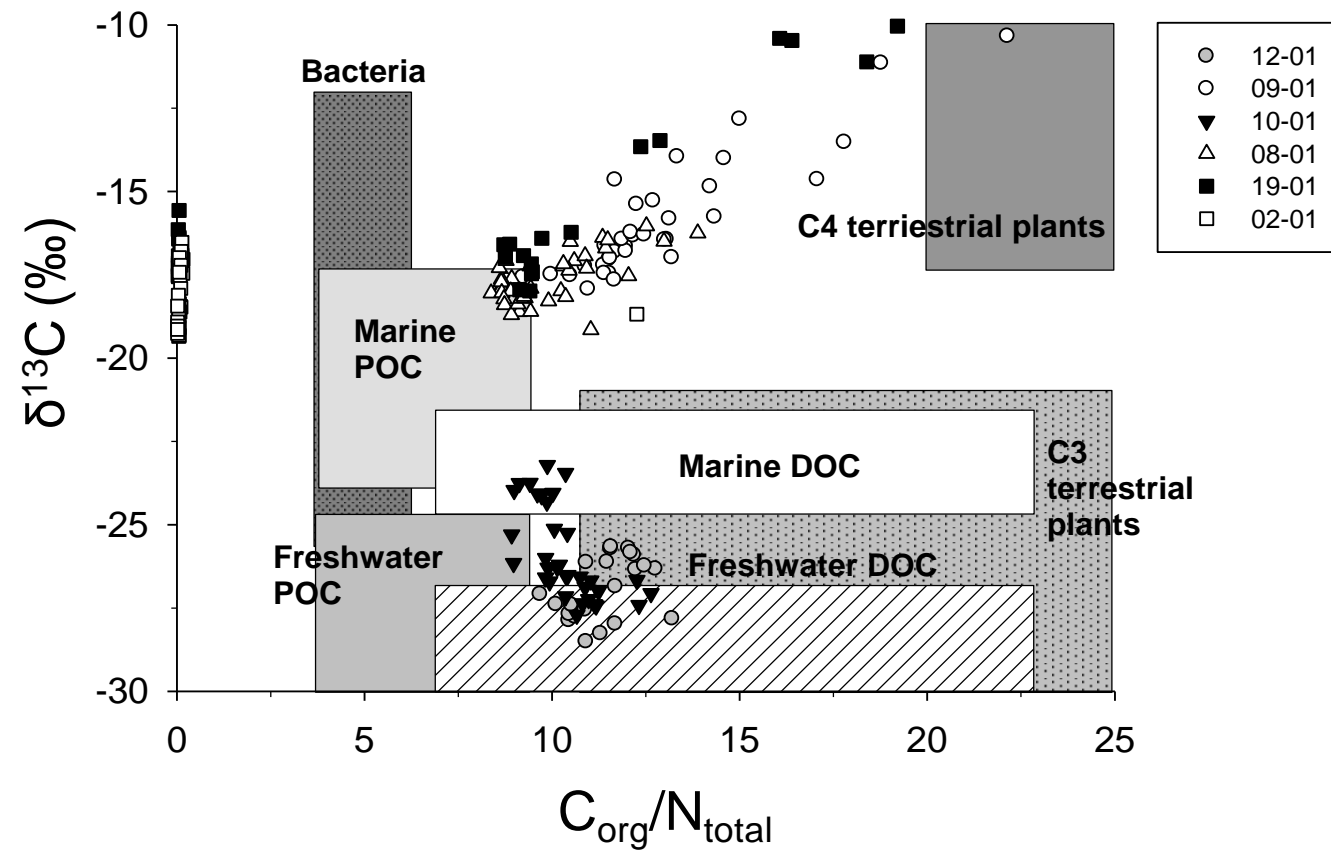
Fig. 8. Carbon stable isotopes $\delta^{13}\text{C}$ (‰) of surface sediments versus salinities of each sampling stations within the Río de la Plata estuary at both sampling dates: grey filled circle lower reaches, black triangles middle reaches, and open circle upper reaches.

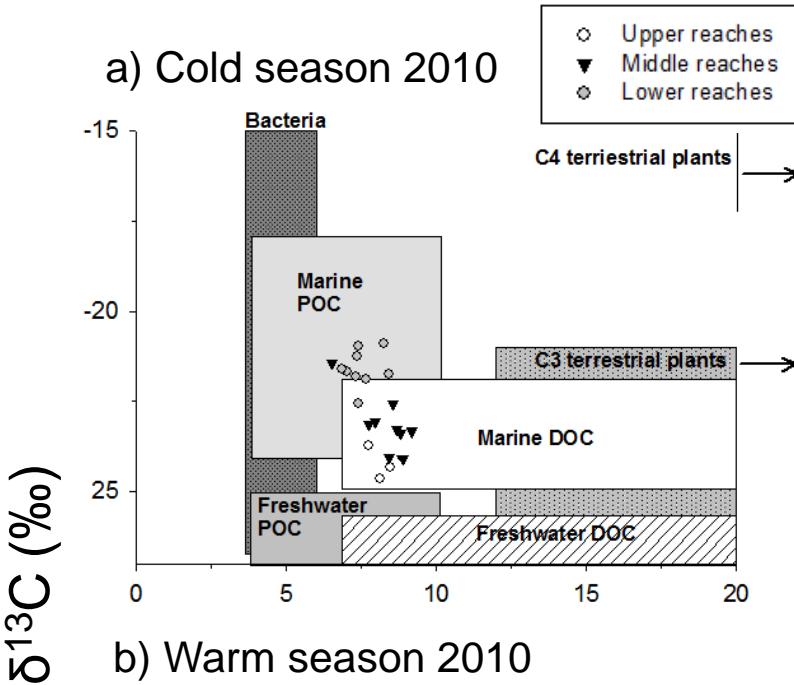




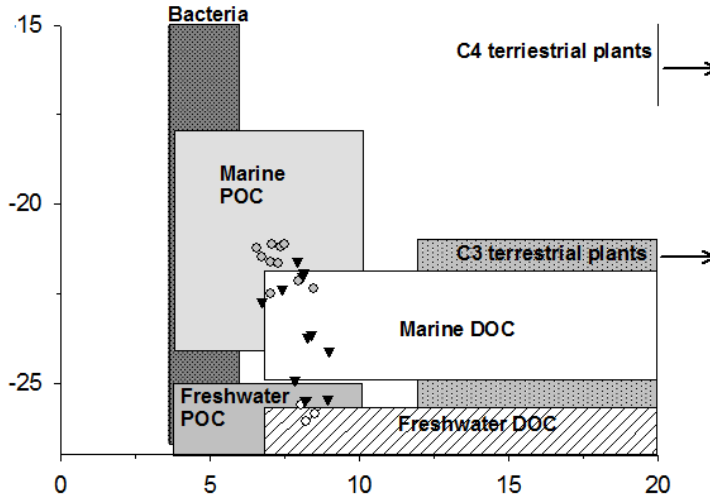




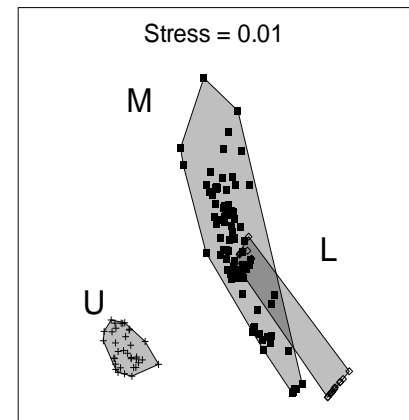




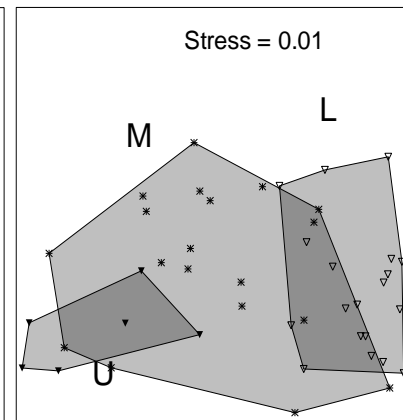
b) Warm season 2010



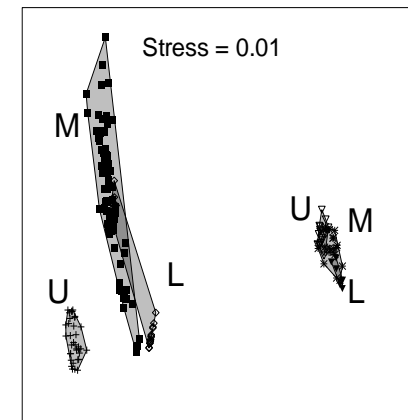
a) Marsh sites

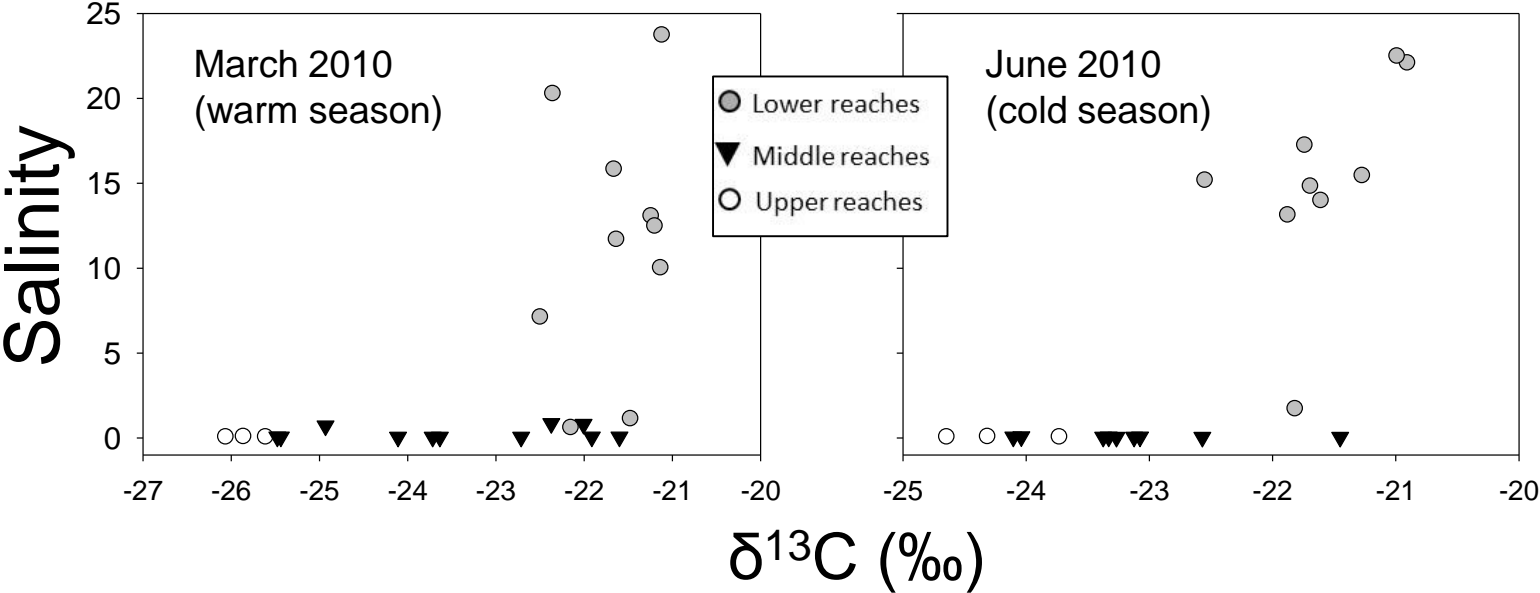


b) Estuarine sites



c) Marsh and estuarine





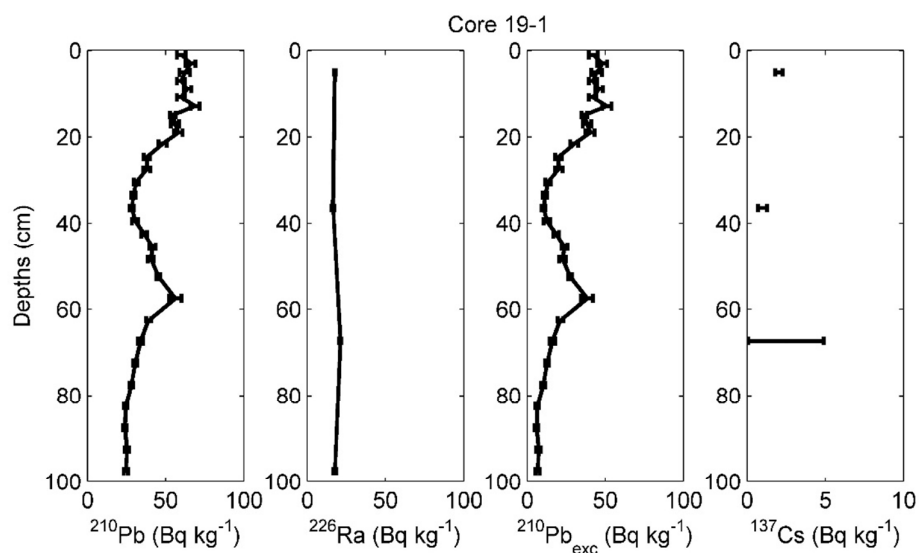


Figure S1: Raw data of radionuclide measurements in core 19-1.

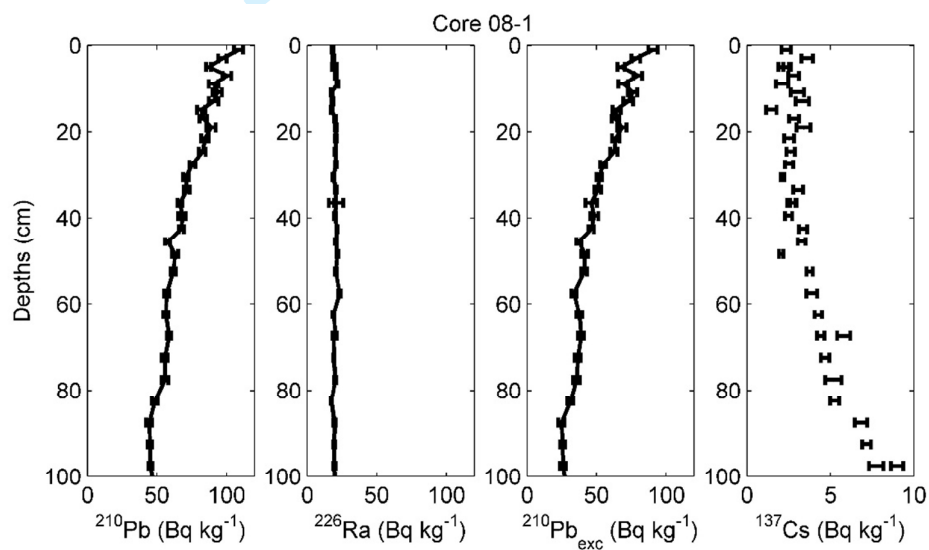


Figure S2: Raw data of radionuclide measurements in core 08-1.

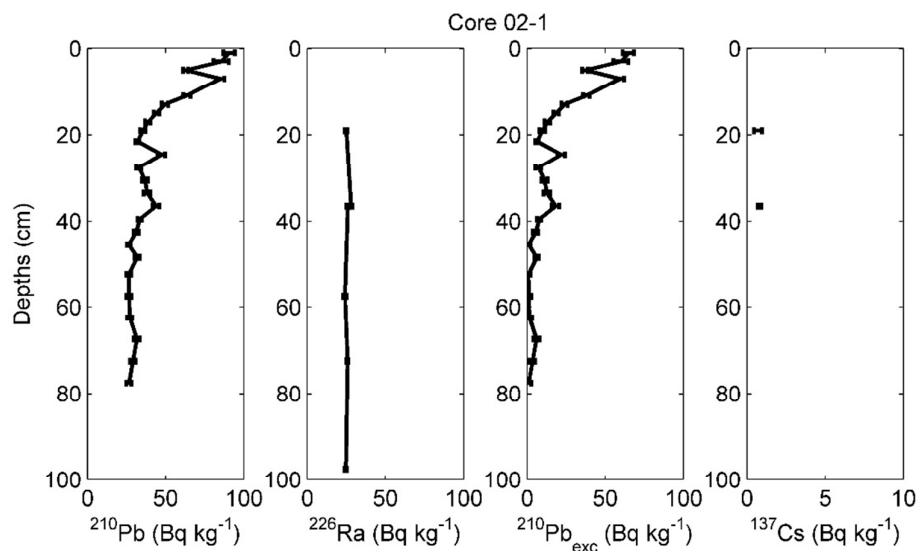


Figure S3: Raw data of radionuclide measurements in core 02-1.

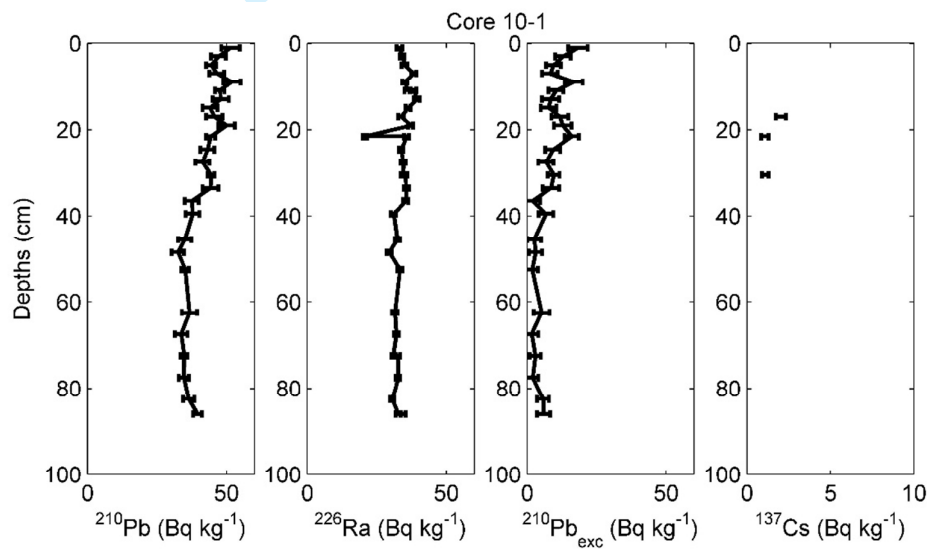


Figure S4: Raw data of radionuclide measurements in core 10-1.

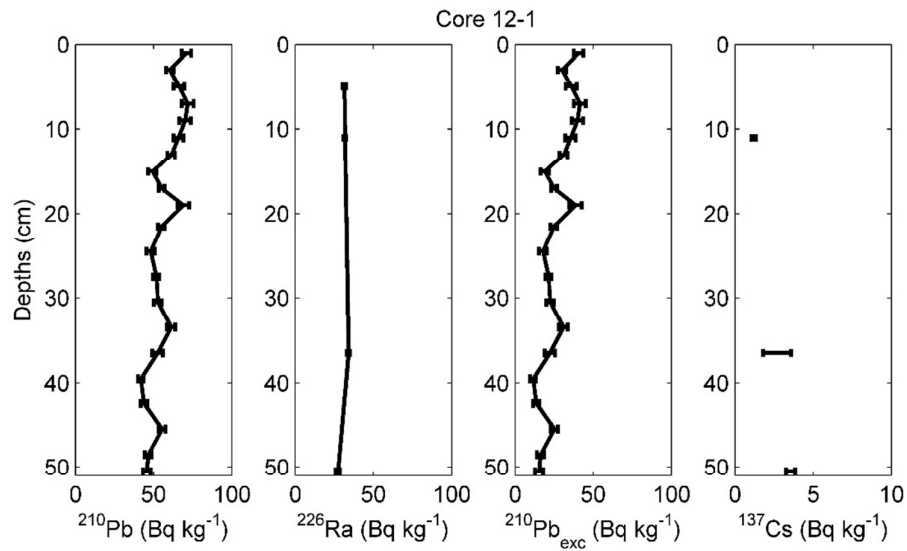


Figure S5: Raw data of radionuclide measurements in core 12-1.

